Smith et al. (1969) did not focus on sulfate formation kinetics; instead, they illustrated through a novel experiment the ability of solid particles to adsorb SO<sub>2</sub> and to release SO<sub>2</sub> during passage through a tube with a wall that adsorbes SO<sub>2</sub>. They measured the number of SO<sub>2</sub> monolayers absorbed on suspended Fe<sub>3</sub>O<sub>4</sub> as function of SO<sub>2</sub> partial pressure. (The monolayer coverage data reported in their Table I are in error by a factor of 100 too large; e.g., the number of monolayers at 1.13 ppm should be 0.38 x  $10^{-2}$ .)

Chun and Quon (1973) measured the reactivity of ferric oxide to SO<sub>2</sub>, using a flow system involving a filter containing suspended particles. They determined a removal rate constant of 9.4 x  $10^{-3}$  ppm<sup>-1</sup> min<sup>-1</sup> [-d(ln  $\rho$ )dt], where  $\rho$  is the fraction of surface sites available for reaction. Extrapolating this to an atmospheric particle concentration of 100 ug/M<sup>3</sup> with an equivalent reactivity and an SO<sub>2</sub> concentration of 0.1 ppm, the data project an atmospheric removal rate of 0.1 percent per hour.

Stevens et. al. (1978) report total iron concentrations in six U. S. cities ranging between 0.5 and 1.3  $ug/m^3$ . Other species such as manganese, copper, or vanadium had total concentrations usually below 0.1  $ug/m^3$ . Thus actual ambient air concentrations are a factor of approximately 50 times less than that assumed by the authors in the above papers. A reactive particle concentration of 2  $ug/m^3$  would yield a predicted SO<sub>2</sub> removal rate of no more than 0.002 percent fieari

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